

NIOH and NIOSH basis for an occupational health standard:  
Di(2-ethylhexyl)phthalate (DEHP)

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## PREFACE

A memorandum of understanding has been signed by two government agencies in the United States and Sweden - the Division of Standards Development and Technology Transfer of the National Institute for Occupational Safety and Health, U.S. Department of Health and Human Services (DSDTT/NIOSH), and the Criteria Group, the National Institute of Occupational Health (NIOH). The purpose of the memorandum is to exchange information and expertise in the area of occupational safety and health. One product of this agreement is the development of documents to provide the scientific basis for establishing recommended occupational exposure limits. The exposure limits will be developed separately by the two countries according to their different national policies.

This document on the health effects of occupational exposure to di(2-ethylhexyl)phthalate (DEHP) is the second product of that agreement. The document was written by Per Garberg, Johan Högberg, Ingvar Lundberg, and Per Lundberg (NIOH) and was reviewed by the Criteria Group and by DSDTT/NIOSH.

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#### Abbreviations

DEHP	Di(2-ethylhexyl)phthalate
MEHP	Mono(2-ethylhexyl)phthalate
IPCS	International Programme on Chemical Safety
i.p.	intraperitoneal
i.v.	intravenous
p.o.	per os
HPLC	high pressure liquid chromatography
TLC	thin-layer chromatography

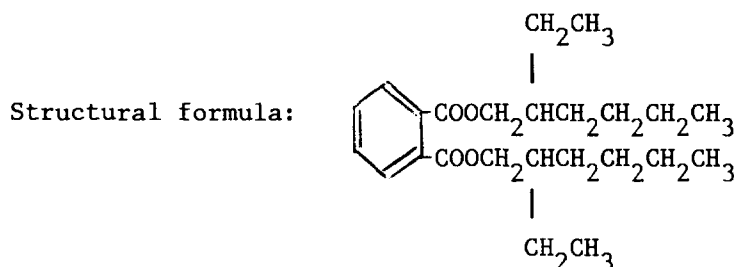
## 1 INTRODUCTION

(The literature is followed to August, 1987.)

Diethylhexylphthalate (DEHP), a diester of phthalic acid, is used primarily as a softener in plastics, primarily for polyvinyl chloride (PVC). Smaller amounts are used as lubricants. Occupational exposure occurs during the production of DEHP, during the addition of DEHP to plastics, and in the processing of softened plastics.

## 2 IDENTITY, PHYSICAL AND CHEMICAL PROPERTIES

The facts and figures in this section are taken from the following references: 23, 33, 45, 49, 53, 66, 79, 88, 102, 113, 121, 122, 125



Molecular formula: C<sub>24</sub>H<sub>38</sub>O<sub>4</sub>

Molecular weight: 390.62

Common synonyms: 1,2-Benzenedicarboxylic acid bis(2-ethylhexyl) ester (CAS name); Phthalic acid bis(2-ethylhexyl) ester (IUPAC name); BEHP; 1,2-Benzenedicarboxylic acid bis(ethylhexyl) ester; bis(2-ethylhexyl) 1,2-benzenedicarboxylate; bis(2-ethylhexyl) ester of phthalic acid; bis(2-ethylhexyl) phthalate; di(2-ethylhexyl) ortho-phthalate; di(ethylhexyl) phthalate; DEHP; dioctyl phthalate; DOP; ethylhexyl phthalate; 2-ethylhexyl phthalate; octyl phthalate; phthalic acid dioctyl ester.

CAS registry number: 117-81-7.

Conversion factors:  $1 \text{ ppm} = 15.87 \text{ mg/m}^3$ ;  $1 \text{ mg/m}^3 = 0.063 \text{ ppm}$ .

Melting point: - 46 °C (pour-point)

Boiling point: 370 °C (101.3 kPa)

236 °C (1.33 kPa)

231 °C (0.67 kPa)

Flash point: 425 °C (open cup)

Density: 0.98 g/ml (20 °C)

Vapour pressure: 0.0013 kPa

DEHP is a colourless to yellow, oily liquid at room temperature and normal atmospheric pressure. DEHP is available in a variety of technical grades.

The solubility of DEHP in water is low, 0.3-0.4 mg/l at 20-25 °C, it is much less soluble in salt water. DEHP is miscible with most common organic solvents and it is more soluble in blood than water. DEHP is lipophilic and the distribution ratio in dichloromethane - Krebs bicarbonate buffer is 1130. The log n-octanol-water partition coefficient is 3-4.

### 3 OCCUPATIONAL EXPOSURE TO DEHP DURING MANUFACTURE, FORMULATION, OR USE

Few data on occupational exposure to DEHP have been reported.

In a phthalate manufacturing industry where DEHP was produced from phthalic anhydride and alcohols, Liss et al (56) measured 8-h time-weighted average workplace air concentrations of DEHP of 0.02-4.1 mg/m<sup>3</sup> among 6 heavily exposed workers. Forty-four other workers in the same plant had exposures below the limit of detection.

In an Italian factory for production of n-butyl phthalate, isobutyl phthalate, and DEHP, Gilioli et al (36) measured total phthalate exposure concentrations between 1 and 60 mg/m<sup>3</sup> with an average air concentration (not stated whether arithmetic or geometric) around 5 mg/m<sup>3</sup>.

Nielsen et al (64) measured total phthalic acid esters in air in a PVC-processing industry in Sweden where diisodecyl phthalate, DEHP, and some butylbenzylphthalate were used. Concentrations between 0.01 and 2.0 mg/m<sup>3</sup> were recorded in 96 2-hour personal samples from 54 workers.

Air concentrations of total phthalates were measured in a Russian PVC-processing plant using mainly dibutyl phthalate and higher alkyl phthalates but also some DEHP and other phthalates. Concentrations between 1.7 and 66 mg/m<sup>3</sup> were recorded (59).

DEHP air concentrations between 0.09 and 0.16 mg/m<sup>3</sup> were reported from a German factory for phthalate production (110).

#### 4 KINETICS AND METABOLISM

##### 4.1 Absorption

Experiments with radioactive (<sup>14</sup>C-labeled) DEHP have indicated that at least 50% of the radioactivity of a single oral dose (2.9 mg/kg) is absorbed in the rat intestine since 42 % and 14 % were excreted in urine and in bile, respectively (27). The same authors also found that DEHP was rapidly hydrolyzed by pancreatic lipase, indicating the possibility that DEHP is hydrolyzed in the gut before absorption. DEHP analysis in liver further supported this supposition in that no unmetabolized DEHP was found in liver after administration of smaller (<0.4 g/kg) oral doses. At higher doses (>0.5 g/kg) the analyses were positive (8). At a dose of 2 g/kg given p.o. to rats the bioavailability of DEHP was 14% as measured with a HPLC technique in blood. When a higher dose (4 g/kg) was given i.p., only 5% was recovered in blood, again indicating a role for hydrolysis in the gut in the kinetics of DEHP (77). Experiments with p.o. administration of MEHP indicated that this metabolite was well absorbed (109). Radioactive MEHP or



DEHP was given to rats and 16 times more radioactivity from MEHP than from DEHP was recovered in plasma.

P.o. administration of DEHP to young rats (1 g/kg) lead to a larger MEHP area under the plasma concentration time curve than in older rats when measured with gas chromatography (103), indicating either a more rapid hydrolysis of DEHP or a more efficient absorption of MEHP in young rats.

The systemic availability of DEHP is low (5% when 4 g/kg was given) after i.p. administration. Relatively low amounts of MEHP were recovered in blood in these experiments (77).

#### 4.2 Distribution

Intravenously (i.v.) administered DEHP is rapidly eliminated from blood. This has been demonstrated in experiments where radioactive DEHP was injected i.v. and blood levels were followed by TLC (92). At a low dose (0.1 mg/kg) there was an initial phase with a half-time of 4.5 min and a second phase with a half-time of 22 min. At a higher dose (200 mg/kg) the initial phase had a half-time of 9 min. This indicated that DEHP was taken up in a tissue compartment by a saturable process (92). Radioactivity from DEHP was rapidly distributed to liver, carcass, lung and spleen when administered i.v. (27, 92).

Perorally administered DEHP is mainly distributed as MEHP (77, 109). In the liver unmetabolized DEHP was recovered only after high doses (>0.5 g/kg) of DEHP, indicating a threshold phenomenon in its absorption and distribution (1, 8). Several reports indicate that DEHP is more or less completely hydrolyzed in the gut (8, 27, 77, 109). The disposition kinetics of MEHP have been analyzed by Pollack et al (77) and by Teirlynck and Belpaire (109) who found that the peak concentration of MEHP in blood was reached 15 min after p.o. or i.p. administration of MEHP (77). The half time of MEHP in blood or plasma is shorter than for DEHP (77, 109). The in vitro plasma protein binding of MEHP is extensive (103).

Newborn infants receiving exchange transfusions may reach plasma levels of about 10 µg/ml of DEHP (104). This level is about twice as high as those found in leukaemia patients receiving platelet concentrates and about five

times as high as levels found in haemodialysed patients. After treatment this level rapidly fell to about 3 µg/ml within 2 hours, and then a further drop with a half-time of about 12 hours (104).

#### 4.3 Metabolism

DEHP is hydrolyzed in vitro by pancreatic lipase to MEHP (27), indicating that this metabolism mainly occurs in the gut lumen. In rats about 80% of a p.o. dose of DEHP undergoes mono-de-esterification (77), while intraarterially administered DEHP was only slowly converted to MEHP (77). The hydrolysis in rat liver has been studied with both negative (27) and positive (6) results.

MEHP may be further metabolized in the gut wall (77) or other organs. Rat liver cells in culture as well as the intact rat converts MEHP to several metabolites (7, 55). The  $\omega$ - and  $\omega$ -1-carbon oxidation products constitute more than 85% of the metabolites (7, 55, 60). The ethyl side chain may also be oxidized (55). It has been suggested that  $\omega$ -oxidation leads to a product which is further metabolized by  $\beta$ -oxidation in the peroxisomes (7, 55). A non-linear dose-dependency of this metabolism in the rat has been observed; the dominating role of  $\omega$ -oxidation over  $\omega$ -1-oxidation was increased by high doses of MEHP (55).

Small amounts of radiolabel derived from 2-ethyl-(1-<sup>14</sup>C)-hexyl-labeled DEHP was recovered in a DNA fraction (8). However, the metabolite responsible for this radioactivity or the nature of its binding was not fully explored.

There are marked animal species differences in the metabolism of DEHP. Thus the  $\omega$ -oxidation seems to play a dominate role in the rat and guinea pig (8, 54, 55), but seems to be a minor pathway in the mouse, hamster, green monkey, cynomolgus monkey, and marmoset (8, 54). In urine of DEHP treated rats, no conjugated metabolites have been detected. A minor portion is conjugated in the hamster and a major portion is conjugated in the marmoset, mouse, guinea pig, green monkey and, man (8, 54). Only glucuronides have been found in these species (8).

The amount of unmetabolized MEHP excreted in urine also varies considerably between species; it is low in the rat and hamster, but high in mouse, guinea pig, green monkey and man (8).

Repeated oral administration of DEHP or MEHP at high doses (500 mg/kg) to rats leads to a change in the metabolic profile. There was an increase in  $\omega$ -oxidized metabolites and a decrease in  $\omega$ -1-oxidized metabolites (55). Drug metabolizing enzyme activities have been studied after DEHP administration and in some cases changes have been observed (2, 37, 119).

The same metabolites as those found in rat urine can be detected in human urine. One study on i.v. injected DEHP (8) and one on orally administered DEHP (90) indicated that man metabolizes DEHP by  $\omega$  and  $\omega$ -1-oxidation as well as by oxidation of the ethyl side chain. However, the  $\omega$ -oxidation-pathway seems to be a minor pathway in man. More than half of the metabolites recovered in human urine were conjugated metabolites (8, 90).

Time-averaged concentrations of DEHP, MEHP and phthalic acid in blood, reported for patients undergoing maintenance hemodialysis, were 1.9, 1.3 and 5.2  $\mu\text{g/ml}$  respectively (76). These patients are considered to be at risk of potential DEHP toxicity through prolonged contact periods with medical plastic products that contain DEHP. The relatively high circulating level of phthalic acid may indicate an altered metabolism of DEHP in uremic patients (76).

Levels of DEHP and MEHP in plasma were studied in newborn infants receiving blood exchange transfusions. In one case the MEHP half-life was the same as for DEHP (about 12 hours), indicating that the hydrolysis of DEHP was the rate limiting metabolic step. However, in other children the half-time of MEHP was longer than that of DEHP, suggesting a low oxidizing capacity in most newborns (104).

#### 4.4 Excretion

Radioactivity from intravenously injected  $^{14}\text{C}$ -labeled DEHP is mainly recovered in urine and feces after 24 hours (92), indicating that urine and bile are major excretory pathways. At a low dose (0.1 mg/kg) given to rats, 50-60% of injected radioactivity was recovered in urine and feces after 24

hours, while less than 50% was recovered at a high dose of 200 mg/kg (92). Seven days after an oral dose (2.9 mg/kg) of DEHP, 42% of the radioactivity was recovered in urine and 57% in feces (27). In these experiments on rats the biliary excretion was also quantitated, and it was found that 14% of the radioactivity was recovered in bile after 4 days (27). The almost 100% recovery documented by Daniel & Bratt (27) has been confirmed by Teirlynck & Belpaire (109). Oral administration of MEHP (50-500 mg/kg) gave a higher urinary recovery than orally administered DEHP (50-500 mg/kg) as measured after 24 hours (55).

Urinary excretion in man has been quantitated after oral administration of non-radioactive DEHP (0.45 mg/kg). It was found that 15-25% was excreted in urine as MEHP or oxidized metabolites within 2-3 days (90).

In the rat no unmetabolized DEHP is excreted in urine, while small amounts are found in mouse or green monkey urine (8). Major amounts of MEHP are excreted in the urine by mouse, guinea pig, green monkey and man (8). However, oxidized metabolites constitutes a major portion of excretion products in urine from rat, mouse, hamster, green monkey, and man (8).

Changes in excretion pathways have been observed after prolonged administration of DEHP. In control rats the fecal excretion pathway dominated after p.o. administration, while in habituated rats (7 days feeding with DEHP) the urinary pathway dominated (27).

#### 4.5 Retention and turnover

At least two elimination phases of radioactivity, given as DEHP, have been observed in rat blood, but these have short half-lives (4.5-9 and 22 min respectively) (92). In the liver after 7 weeks of oral administration the elimination phase was considerably slower, with a half-life of 3-5 days (27). No accumulation of DEHP or MEHP was observed after dosing with 2.8 g/kg/day for 7 days (109). No accumulation was reported from a long term feeding study at doses of 1000 or 5000 ppm in the food (corresponding to a daily dose of about 50 and 250 mg/kg) for 5-7 weeks (27).

Analysis of the total amount of urinary metabolites, measured as derivatized phthalic acid, indicated a weak positive correlation between

occupational exposure to phthalate and the metabolites (56, 64). In the first study (64) workers were mainly exposed to DEHP and diisodecyl phthalate. In the second study (56) workers were exposed to DEHP and phthalic anhydride. The authors concluded that phthalic anhydride influenced the urinary level more than did DEHP.

#### 4.6 Summary

DEHP is hydrolyzed in the gut. In the liver unhydrolyzed DEHP can only be found after administration of high doses. The hydrolyzed metabolite MEHP is rapidly absorbed and is further metabolized in the liver. In the rat the  $\omega$  - and  $\omega$  -1-carbon oxidation products constitute more than 85 % of the metabolites. It has been suggested that oxidation products are further metabolized by  $\beta$ -oxidation in the peroxisomes. This metabolism is dose- and frequency of administration-dependent. The metabolism is changed by repeated administration of DEHP. There are great species variations, and the  $\omega$  -oxidation pathway seems to play a minor role in man. Major excretory pathways are urine and bile. In feces and urine almost 100 % of an administered dose can be recovered. No accumulation of DEHP or MEHP has been observed.

### 5 EFFECTS ON EXPERIMENTAL ANIMALS AND IN VITRO TEST SYSTEMS

#### 5.1 Single Exposure

Numerous LD<sub>50</sub> values have been reported for DEHP. The oral LD<sub>50</sub> values in rats generally exceeded 25 g/kg and in mice, 30 g/kg (66, 125). The oral LD<sub>50</sub> value for the rabbit is given as 33.9 g/kg (95), and for the guinea pig, 26.3 g/kg (49).

After i.p. administration to rats an LD<sub>50</sub> value of 30.7 g/kg has been reported (95). In mice the value is noted to be 37.8 g/kg (51). In other studies LD<sub>50</sub> values in the range 14-75 g/kg have been calculated for the mouse after i.p. administration (125).

LD<sub>50</sub> values in the range 200-250 mg/kg were reported for the rat after i.v. doses of DEHP solubilized in nonionic detergent (84, 91).

From the information provided in the acute studies the main sign of toxicity after single oral or intraperitoneal dosing is diarrhoea, possibly due to the oily nature of the ester. An i.p. dose of 500 mg/kg in rats decreased spontaneous running activity indicating behavioural changes (85). After i.v. dosing, lung lesions characterized by oedema, haemorrhage, and infiltrations of polymorphonuclear leucocytes have been observed in the rat (93). The effects were evident at doses as low as 50 mg/kg. The mechanism of the aetiology of the lung lesions is unknown. It has, however, been suggested that some changes could be due to the release of lysosomal enzymes from alveolar macrophages, which has been seen in vitro in rabbit tissues treated with DEHP (11).

Treatment i.v. of rabbits with DEHP induced a fall in blood pressure and an increase in breathing rate at a dose of 350 mg/kg. No deaths occurred after doses up to 650 mg/kg (20).

The monoester, MEHP, may be more toxic than the diester but data are very limited. In a short note (118) the oral LD<sub>50</sub> in the rat was reported to be 1.34 g/kg (female) and 1.8 g/kg (male).

## 5.2 Short-Term Exposures

The most obvious sign of toxicity in the rat after oral administration is a reduction in the rate of weight gain, usually evident within a few days of dosing of 0.4 g/kg/day or above (40, 65). Reduction in body weights were observed in rats given dietary levels of 12,500 and 25,000 ppm for 13 weeks. Dosages of 1600-6300 ppm resulted in either slight elevations of body weight or had no effect (62).

Reductions in haemoglobin, packed cell volume and erythrocytes have been observed in rats given 1% or 2% dietary DEHP for 17 weeks but not at 0.2% for the same period (40).

Doses of more than 3 g/kg/day for periods up to 90 days were necessary to cause deaths in rats (65, 95). No deaths occurred in rats after dietary dosing for 14 days at 50,000 ppm or below (62).

MEHP also reduced the body weight gain when given in the diet of rats at 6,400 ppm (22). No effects on body weight occurred at dietary levels up to 625 ppm for 3 months. At 625 ppm a significant reduction in blood glucose was observed.

Decreases in body weight have also been observed in mice given dietary DEHP for 14 days at levels from 6,300 ppm to 100,000, or for 13 weeks at 1000 ppm to 12,500 ppm (62). Cystic kidneys and centrilobular necrosis were noted in one strain of mice (ddY) fed 0.25 or 2.5% DEHP for 2 weeks but not in another study using a different strain (B6C3F<sub>1</sub>), higher exposure levels, and longer periods (125).

DEHP administered i.v. at 25-500 mg/kg/day for 2-4 weeks to beagle dogs resulted in pulmonary haemorrhage and inflammatory response similar in appearance to the "shock-lung" effect (125).

A few inhalation studies have been reported. However, due to the omission of experimental details they can not be assessed.

### 5.3 Biochemical effects

A large amount of research work has been carried out on the hepatic effects such as cell proliferation, of DEHP in laboratory animals. A single dose of DEHP, 500 mg/kg, resulted in a six-fold increase of male rat hepatocytes in S-phase and a 15-fold increase of male mouse hepatocytes in S-phase 24 h after dosing. Three weeks of feeding DEHP at 6000 ppm in the diet increased liver/body weight ratios to approximately 140% that of controls in female rats and 170% of controls in male mice (19). However, at the end of 2 years of feeding DEHP at 0.03, 0.1 or 1.2% in the diet there was no detectable liver cell proliferation as studied by <sup>3</sup>H-thymidine incorporation (78). In vitro studies also have shown that DEHP decreases the protein kinase c activity (100).

The cell proliferation induced by DEHP forms the basis of one theory of DEHP hepatocarcinogenesis. This theory argues that the increased cellular growth induced by DEHP may result in spontaneous mutational events or promotional effects (16, 19).

Peroxisomes (microbodies) are small cytoplasmatic organelles. Functionally, peroxisomes have been implicated in gluconeogenesis, detoxification of  $H_2O_2$  and in lipid metabolism (82). Some chemical agents may cause peroxisome proliferation in mouse and rat liver. This property was first discovered for some hypolipidemic drugs, such as clofibrate and naphenopin, but has later been found to be common to a structurally diverse group of chemicals including DEHP. Peroxisome proliferators including DEHP and other phthalates are liver carcinogens, when administered in the diet (80).

There seems to be a fair correlation between the ability of a chemical to induce peroxisome proliferation and its liver carcinogenic properties (81).

According to the peroxisome proliferation hypothesis for liver carcinogenesis, DEHP and/or its metabolites reduces lipid synthesis and oxidation so that fatty acid derivatives (e.g. triglycerides) are accumulated in the hepatocyte. Genes for lipid metabolizing enzymes, including peroxisomal beta-oxidation enzymes, are then expressed. It has recently been shown that MEHP and the  $\omega$ -1-carbon oxidation products are the DEHP metabolites responsible for peroxisome proliferation in primary hepatocyte cultures (60).

The increased lipid metabolism in the peroxisomes may generate increased amounts of  $H_2O_2$ . Since catalase, the enzyme which catalyzes the breakdown of  $H_2O_2$  to water, is only slightly induced by DEHP the result will be an increased formation of oxygen radicals which may damage membranes and/or DNA (16). If the exposure is prolonged a series of cellular insults may result in permanently altered cells which form foci and ultimately tumors. It should, however, be clear that this hypothesis has only limited experimental support.

DEHP is a moderately effective inducer of peroxisome proliferation in the rat. Chronic administration of doses of 50 mg/kg/day seems to be effective as measured by an increase in peroxisomal enzymes (61). However, doses of 25 mg/kg/day seem not to increase these enzyme activities in the rat liver (50).

Another experiment has shown a large increase of hepatic peroxisomes in the rat after 14 days of oral DEHP administration at 2000 mg/kg/day. When the



same dose was given to marmosets there was no peroxisome proliferation. Moreover, peroxisomal enzymes showed a pronounced rise in the rat but not in the marmoset (83). The difference may have been caused not only by a less efficient hydrolysis of DEHP and therefore less efficient absorption in the marmoset (83) but also by different intrinsic responsiveness to MEHP by the rat and marmoset hepatocytes (60). There is some experimental support for both of these explanations. Some experiments suggest that human hepatocytes may be similar to marmoset hepatocytes in their peroxisomal response to MEHP (16).

Nothing is known about rat, marmoset, or human peroxisome responses to DEHP by inhalation exposure.

In vitro, rat liver microsomal cytochrome P-450 levels were not affected by DEHP (37). However, when rats were treated in vivo with DEHP by i.p. injections (119) or repeatedly given DEHP orally (2), an increase in cytochrome P-450 levels has been observed.

In the liver, mitochondrial enzymes and mitochondrial morphology have been affected by DEHP administration (74, 96). The significance of the effects on mitochondria is not known as the mitochondrial functions usually appear to be normal (125).

In male rats given DEHP in the diet, urinary excretion of Zn is enhanced and the testicular levels of Zn decreased (42, 68). See also 6.

A number of studies has shown that DEHP given in the diet (0.5-2%) causes lowering of triglyceride and cholesterol levels in the plasma of the rat (13, 14, 15, 24, 87, 127, 128). DEHP inhibits the biosynthesis of cholesterol, an effect which is accompanied by phospholipidosis. The same effects are seen with MEHP (72).

#### 5.4 Skin and eye irritation; sensitization

The data available indicate that DEHP is a weak irritant to mammalian skin on topical or intradermal exposures (20, 125).

In one study no irritation in the rabbit eye occurred when DEHP was instilled into the eye (51).

There are no data indicating that DEHP should be a sensitizing agent.

#### 5.5 Long-Term Exposures

In a 24 month study (44) groups of 43 male and 43 female Wistar rats were fed diets containing 0, 0.1 or 0.5% DEHP, with interim kills at 3, 6 and 12 months. Because of high mortality only two control, four low dose and seven high dose animals were alive at termination. During the first year DEHP slightly reduced body weights but by the second year the weights of all groups were similar. During the first six months an increase in relative liver and kidney weights were seen in DEHP treated animals but later they were similar in control and treated animals. After 3 months 1/8 rats in the low dose group were found to have mild tubular atrophy in the kidney. After 6, 12 and 24 months no compound related pathology was evident.

In another 24 month study (21) groups of 32 male and 32 female Sherman rats were given diets containing 0, 0.04, 0.13 or 0.4% DEHP. Because of reduced life-expectancy of the animals and the small numbers of animals used, the study is inadequate to assess the chronic toxicity of DEHP.

In a 12 month study (65) a group of 20 male and 20 female Wistar rats were given a diet containing 0.35% DEHP. A control group received the diet only. The only gross or micropathology noted in dosed animals at necropsy was hepatomegaly. During the study, however, about 30% of the animals died due to congestion of the small intestine and loss of the gastric or intestinal mucosa, which was complicated by purulent pneumonia and endometritis.

In a 2 year study (47, 62) groups of F344 rats were given dietary levels of 0, 6000 or 12000 ppm DEHP. A decreased body weight in exposed groups was noted from around week 30 until termination. Except for neoplastic effects (section 8) and testicular atrophy (section 6) a compound related hypertrophy of cells of the anterior pituitary were noted at necropsy of the high dose rats. In both of the exposed groups an increased incidence of clear cell changes in the livers was observed.

In the same study (47, 62) B6C3F<sub>1</sub> mice were exposed to 3000 or 6000 ppm DEHP. A dose related decrease in dosed female mice body weights was observed. There was no increased incidence of non-neoplastic lesions except for seminiferous tubular degeneration of the testes in male mice (section 6).

In a study with groups of male and female guinea pigs fed diets containing 0, 400 or 1300 ppm DEHP for 12 months (21), dosed animals had statistically higher body weights than controls and a slight increase in liver weights was noted in dosed females. No other compound related lesions were found.

#### 5.6 Summary

Based on LD<sub>50</sub> studies DEHP has a low acute toxicity. A prolonged dosing of DEHP to animals produced proliferation of liver peroxisomes, hepatomegaly and a reduction of body weight gain.

DEHP is a weak irritant to mammalian skin but there are no data indicating that DEHP should act as a sensitizing agent.

### 6 REPRODUCTIVE, EMBRYOTOXIC, AND TERATOGENIC EFFECTS

#### 6.1 Reproductive effects

The effects of DEHP on male reproductive organs have been studied extensively. The majority of the studies have been carried out using rats or mice given DEHP in the diet.

Seminiferous tubular atrophy, comprising a loss of spermatids and spermatocytes, was the result when 4-week-old Wistar rats were administered 2800 mg DEHP/kg by oral intubation for 10 days (39). In similarly treated 10-week-old rats about 50% of the tubules were atrophic and the remainder unaffected. No testicular damage was seen in treated 15-week-old rats. When DEHP was given in the diet (2%; approximately 1200 mg DEHP/kg) daily to 4-week-old rats, the lesions produced were reversible whether treatment was stopped prior to, or continued until after the control rats had reached sexual maturity.

In rats given 1 or 2% dietary DEHP the effect was dose-dependent with visible signs of atrophy after approximately 2 weeks of feeding (40). The atrophy was accompanied by pituitary changes, namely enlargement and vacuolization of the basophils of the pars distalis corresponding to the formation of the so-called castration cells seen after gonadectomy (40). Similarly, there was a concomitant reduction in testicular and prostatic zinc levels with increased urinary excretion of zinc (42).

In a rat study with orally administered DEHP (2% in the diet) the serum testosterone levels were reduced by approximately 50% but the levels in testes rose by 150% (70). Simultaneous administration of testosterone or zinc had no protective effect on the atrophy but did prevent the weight-reduction of the accessory sex organs (39, 73).

DEHP in the diet (2%) given to 7 young Wistar rats for one week decreased the testicular weight significantly ( $p < 0.05$ ) compared to controls (67). The testicular zinc concentration was decreased as well. In another study (68) groups of 20 male rats were given 2.0 g/kg/day by gavage for 14 days. Ten rats were then killed and the remaining 10 rats were followed for an additional 45 day period without administration of DEHP. The histopathological changes of the testes seen at day 15 were: a marked shrinkage of the seminiferous tubules; germinal epithelium consisting only of Sertoli cells; very few spermatogonia. After 45 days the percentage of spermatogenic tubules had increased from 0 (at day 15) to 12.8%, indicating a limited reversibility of the testicular atrophy.

Similar degeneration of the seminiferous tubules were seen (89) when 13-week-old Wistar rats were given 2 g DEHP/kg p.o. for 7 consecutive days.

The changes induced in rats by DEHP may be age dependent (106). Younger animals appeared to be more susceptible to testicular injury and were slower to recover (39). If treatment was withdrawn before the rats were sexually mature the effects were reversible (39). Plasma levels of DEHP were similar in sexually mature and immature rats indicating no difference in absorption (107).

Rats given i.p. injections of DEHP did not show any strong evidence for testicular atrophy (25). As with oral dosing, however, i.p. administration

of DEHP decreased serum testosterone levels (69). Both routes have also altered the activities of some testicular enzymes (25, 94).

Some degenerated primary spermatocytes and altered Sertoli cells were observed in Sprague-Dawley rats given 3 h infusions of an emulsion at a rate of 1 ml/h corresponding to a daily dose of 500 mg DEHP/kg (103). The infusions were given every other day on six occasions. The emulsion contained DEHP, fractionated egg yolk phosphatides, glycerol, and water. No effects were observed when emulsions corresponding to 0, 5, or 50 mg DEHP/kg were given.

In another study (26) groups of Swiss-Webster mice were given daily i.p. injections of 50 or 100 mg DEHP/kg for 5 days or alternate daily i.p. injections of 50 or 100 mg DEHP/kg for 20 days (10 injections). The animals were killed 24 h following the last injection. No significant alterations in testicular weight or zinc levels occurred. In Sprague-Dawley rats given similar treatment the gonadal zinc level decreased about 30%.

In a NTP-sponsored study (58) with CD-1 mice given 0.3% DEHP in the diet, the animals had significantly diminished testis and epididymis weights compared to controls. In addition, the sperm concentration in the cauda epididymis was decreased and the percent abnormal sperm in the cauda was significantly higher in the treated mice.

A high oral dose (4.2 g/kg) of DEHP gave a minimal tubular atrophy in hamsters but did not produce any effects on urinary zinc excretion, testicular zinc levels or on testicular weights (42).

The effects of the monoester (MEHP) has not been as well studied. An oral dose of 1 g/kg/day for 5 days to rats produced a significant decrease in testes weights and an extensive testicular atrophy (42). On the other hand, rats given i.p. doses of up to 100 mg/kg of MEHP for 5 days showed no abnormal histology (25) and i.p. doses of 50 mg/kg on alternate days for 20 days produced only a reduction in prostatic zinc levels (26).

In mice fed diets containing 2% MEHP for 1 week (71) testicular zinc and testosterone levels were markedly reduced but there were no reductions in testicular weights. Hamsters given 1 g/kg/day of MEHP for 9 days showed

more severe testicular effects than those given DEHP at levels of 4.2 g/kg/day for the same period (42).

These studies indicate that the rat is the most susceptible species to DEHP-induced testicular atrophy. The mechanism of phthalate-induced testicular damage is not fully understood. The testicular zinc depletion has been suggested as a primary event (34, 35). Zinc is essential for normal testicular function and its depletion is known to lead to testicular atrophy (12). Inhibition of dehydrogenase enzymes, e.g. those controlling the biosynthesis of testosterone, lead to reduced testosterone levels. DEHP administration has been shown to reduce serum testosterone in the rat (69, 70) as well as in the mouse (42), although in the mouse no testicular atrophy was observed. Supplementation with testosterone or zinc did not prevent the testicular damage induced by DEHP (39, 73).

In vitro studies have indicated that the Sertoli cell is the target. Mixed cultures of Sertoli and germ cells prepared from rat testes were exposed to DEHP or MEHP ( $10^{-7}$ - $10^{-4}$ M) (38, 41). DEHP had no effect but MEHP caused a dose-dependent increase in the rate of germ cell detachment from Sertoli cells, accompanied by changes in Sertoli cell morphology.

In another study (58) CD-1 mice were given 0, 0.01, 0.1 or 0.3% DEHP in the diet during a 7 day pre-mating period and a subsequent 98-day cohabitation period. This resulted in complete suppression of fertility in the 0.3% group and a significant reduction in the fertility in the 0.1 group compared to controls. There was no effect of DEHP on fertility in the 0.01% treatment group.

In a fertility study (3) groups of ICR male mice were administered DEHP subcutaneously at dose levels of 1, 2, 5 and 10 ml/kg on day 1, 5 and 10 of the study. At day 21 each male was housed with one virgin female for 7 days. There was a reduction in the incidence of pregnancies in the treated groups (3/8 at the three highest dose levels; 5/8 at the lowest dose level; 14/16 in the control group). On the other hand, repeated matings with virgin females starting on day 2, 6, 11, 16 and 21, and at weekly intervals through 8 weeks, revealed no decrease in the incidence of pregnancy. In the treated groups, examination of pregnant females on day 13 of gestation

revealed an increase in the incidence of preimplantation losses and early fetal deaths.

Only one study (94) has investigated the effects of phthalates on female reproductive organs. Three doses of 4.93 g DEHP/kg at 5-day intervals were given i.p. to female rats. No histopathological changes of the ovaries were seen 22 days after the first injection, but reductions of some enzyme activities were noted. However, i.p. injections of DEHP to male rats did not result in histopathological changes of the testes.

It appears that rats and guinea pigs are sensitive to testicular atrophy caused by DEHP while mice are fairly and hamsters highly resistant. That at least some of the effects associated with this pathology can be produced in vitro argues against a hormonally mediated indirect effect. The earliest effects are seen in Sertoli cells and are described as vacuolation (5).

## 6.2 Embryotoxicity, Teratogenicity

In a study (65) with female Wistar rats given 0.34 or 1.7 g DEHP/kg by gavage during the first 21 days of gestation, the only untoward effects seen were reductions in fetal body weight. Two percent (2%) DEHP in the diet of Fischer 344 rats during the first 20 days of gestation increased the number of resorptions (123).

Intraperitoneal injections of 4.93 and 9.86 g DEHP/kg to Sprague-Dawley rats on days 5, 10, and 15 of gestation resulted in an increased number of resorptions and reduced fetal weights (101). In the highest dose group gross abnormalities, such as twisted hind legs and anophthalmia, were noted but no skeletal defects were observed.

Plasma extracts of two polyvinyl chloride (PVC) plastics containing DEHP were administered intravenously to groups of pregnant Sprague-Dawley rats daily from the 6th to the 15th day of gestation (52). The animals were killed at day 20. The daily doses of DEHP were equivalent to 1.3 mg/kg (low dose) and 5.2 mg/kg (high dose). No significant teratogenic or embryotoxic effects were noted.

Groups of ICR mice were given DEHP in the diet at levels of 0.05 to 1.0% for days 0-18 of gestation whereafter they were sacrificed (97, 99). The food intake was about 7 g/day as an average. At 0.4 and 1.0% levels no live fetuses were found. At the 0.2% level, 40% of the fetuses had malformations including exencephaly, spina bifida and malformed tail. A delayed ossification was seen in about 15% of the fetuses at the 0.1 and 0.2% dose levels.

Dietary administration of 0.025 to 0.15% DEHP on days 0-17 of gestation to CD-1 mice increased the number of resorptions at the 0.1 and 0.15% levels (124).

A single p.o. administration of 0.1 ml/kg DEHP on day 7 of gestation to SPF mice decreased the number and the body weights of living fetuses (116). In another study (126) DEHP was given orally to SPF mice at day 6, 7, 8, 9 or 10 of gestation. When 5.0 or 10.0 ml/kg was given on day 7 there were no live fetuses. 2.5 ml/kg administered the same day produced 14% of living embryos and 1.0 ml/kg gave 40% living embryos. The percentages of live embryos when 10.0 ml/kg was given on day 8, 9 or 10 of gestation were 18, 92 and 95% respectively. Gross and skeletal abnormalities occurred in fetuses given 2.5 and 7.5 ml/kg on day 7 or 8. The abnormalities included exencephaly, open eyelid and club-foot.

In a later study (98) groups of ICR mice were given DEHP by stomach intubation on days 7, 8 and 9 of gestation. The DEHP doses (in olive oil) were 250, 500, 1000, and 2000 mg/kg. The mice were sacrificed on day 18. In the two higher dose groups the numbers of resorptions and malformed fetuses were significantly increased. The malformations most commonly produced were anencephaly and exencephaly. When doses of up to 8000 mg/kg were given i.p. at days 7, 8, and 9 of gestation no effects were noted.

The monoester, MEHP, at oral doses of 225, 450 and 900 mg/kg produced significant signs of maternal toxicity when given to pregnant Wistar rats on days 6-15 of gestation (86). In the highest dose group a 73% mortality was observed in the dams.

In the SPF mouse, oral dosing with 0.1 or 1.0 g/kg MEHP on day 7 of gestation led to increased incidence of early embryonic deaths but dosing



on day 8 or 9 had less effect (117, 126). The fetuses had reduced body weight and a higher incidence of gross abnormalities in fetuses from the high dose group on days 8 and 9 as compared to controls. The mice dosed on day 8 produced fetuses with high incidences of skeletal effects.

Intravenous injections with 11.38 mg/kg MEHP to rabbits on days 6-17 of gestation gave a high incidence of resorptions (114). The incidence of fetal anomalies was similar to controls.

The mechanism of teratogenicity of DEHP or MEHP is not known. Teratogenic activity could result from zinc deficiency, which is known to produce teratogenic effects (108).

In a study (65) with DEHP at doses at 0.34 or 1.7 g/kg/day given to female Wistar rats by gavage for 3 months prior to mating, no effects were seen on the number of resorptions, fetal weights or incidence of skeletal anomalies.

### 6.3 Summary

DEHP can cause testicular damage in rats. The no-effect level is about 0.3 - 0.5 % in the diet. There is evidence that DEHP and its metabolite MEHP are teratogenic and embryo-lethal to rodents. In rats 2 % in the diet throughout the gestation period produced an increased incidence of resorptions and in mice 0.1 % also produced fetal abnormalities. Fertility studies provided contradictory results.

## 7 MUTAGENICITY AND RELATED END-POINTS

The possible genotoxic effect of DEHP has been thoroughly investigated in several different short-term tests. Effects of the major metabolites of DEHP; the monoester (MEHP) and 2-ethylhexanol, as well as phthalic acid and phthalic anhydride, have also been studied.

### 7.1 Mutation

Studies on the possible mutagenic effect of DEHP have been performed in bacteria, fungi and in cultured mammalian cells. *Drosophila melanogaster* has also been used and results from a few in vivo studies on mice have been published.

Many laboratories have used several strains of *Salmonella typhimurium* and doses of DEHP up to 10,000 µg/plate. Incubations both with and without exogenous activation systems have usually been performed. S9-mix from Aroclor 1254 induced rats is commonly used, but other species as well as other inducers have also been used to produce metabolic activation systems. With one exception (115) these test results have been negative (see e.g., 4, 46, 130, 131) and in a IPCS study (9) all five laboratories reported negative results. Other bacteria than *S. typhimurium* have also been used. Negative results were obtained with *E. coli* WP2 at doses up to 2000 µg/plate in one laboratory (130).

The major metabolites of DEHP have also been tested for mutagenic activity in bacteria. Doses up to 3333 µg/plate for MEHP and phthalic anhydride (131) and 2000 µg/plate for 2-ethylhexanol and phthalic acid (4) were used with negative results in strains of *Salmonella* (see also 46, 130). However, one laboratory (115) reported a significant increase in revertants in TA 100 for both DEHP and MEHP, with and without S9, respectively.

The same investigators (115) also detected a dose-dependent (400 and 500 µg/plate) DNA damaging effect to *B. subtilis* in the Rec-assay for MEHP, while DEHP, phthalic acid and 2-ethyl-hexanol were all negative. In this study MEHP showed a positive response in *E. coli* WP2 b/r as well.

Negative results were obtained when pooled urine from rats, treated with DEHP 2000 mg/kg/day for 15 days, were tested for genotoxic activity. A direct plating procedure was used with *S. typhimurium* strains TA-98, TA-100, TA-1535, TA-1537 and TA-1538, both with and without S9 and β-glucuronidase/aryl sulfatase as activation system. When 2-ethylhexanol was tested according to the same protocol was also negative (30).

The induction of mutations by DEHP has been studied in different species of fungi. In the IPCS collaborative study on in vitro assay systems (9) DEHP was considered to be negative in six out of seven assays. Positive results were obtained with *Saccharomyces cerevisiae* both with and without S9 activation at the lowest effective concentrations of 1541 µg/ml and 3081 µg/ml, respectively. Other laboratories using other strains of *S. cerevisiae* or *Schizosaccharomyces pombe* reported negative results at maximum tested doses of 5000 µg/ml.

Mouse lymphoma cells (L5178Y), Chinese hamster V79 cells, as well as human lymphoblasts have been used to study the mutagenic effect of DEHP in cultured mammalian cells. Several investigators have reported negative results, but a few positive results can also be found in the literature.

In the IPCS study (9) only one out of ten investigators reported a positive response. Mouse lymphoma cells were exposed to DEHP, without S9, and two doses were positive (7.5 µg/ml and 20 µg/ml), indicating a positive response for DEHP without S9 in this test. In a separate study (46) where MEHP and 2-ethylhexanol, as well as DEHP were tested in the mouse lymphoma cell assay, all three substances were found to be negative. The doses were 0.016-1.0 µl/ml -S9 and 0.067-5.0 µl/ml +S9 for DEHP and 0.013-1.0 µl/ml for MEHP and 2-ethylhexanol.

DEHP has also been tested for mutagenicity in *Drosophila melanogaster*, using the sex-linked recessive lethal (SLRL) and different somatic recombination and mutation assays. 20 ppm of DEHP administered via injection did not induce SLRLs (129). In the IPCS study (9), however, DEHP gave a positive response in the unstable eye mosaic test in two separate experiments, but only at one dose (6.1 mg/ml), and were weakly active at 611 µg/ml in another eye mosaic assay. No activity however, was seen in the wing spot test at single doses of 6.1 mg/ml. It was concluded that the DEHP test results may be interpreted to show marginal-positive responses (9).

## 7.2 DNA damage

Different end-points, such as unscheduled DNA synthesis (UDS), and single strand breaks have been used to detect damage to DNA induced by DEHP in different mammalian test systems. In a IPCS study (9) negative results were

obtained when single strand breaks were measured, either by alkaline elution in hepatocytes, up to 3907  $\mu\text{g/ml}$  or alkaline sucrose sedimentation in CHO cells, up to 39000  $\mu\text{g/ml}$ . In the IPCS study UDS was also measured in either isolated hepatocytes or cultured HeLa cells by four different laboratories. One investigator detected a positive response of DEHP using isolated hepatocytes. However, since this result was only statistically significant at one dose and not dose-related, the consensus was that DEHP was negative for UDS.

In a separate study DEHP did not induce DNA-repair in primary rat hepatocytes (18). Similarly neither DEHP nor MEHP induced any DNA-repair in primary human hepatocytes, from three different subjects. In this study concentrations as high as 10 mM of DEHP and up to 0.5 mM of MEHP were used and exposure continued for 18 hours. No induction of DNA-repair or increased alkaline elution of DNA were seen in hepatocytes from either female or male F-344 rats treated with DEHP in vivo. Peroxisomal proliferation was induced in male rats treated with 500 mg/kg DEHP at 2, 12, 24 or 48 h before sacrifice, or with 150 mg/kg/day DEHP for 14 consecutive days. Female rats were treated with either 12000 ppm in the diet for 30 days, or 500 mg/kg 2 h before sacrifice, or the combination of both treatments.

Similar in vivo/in vitro measurements of unscheduled DNA synthesis (UDS) in hepatocytes were reported by Kornbrust et al (48). Unscheduled DNA synthesis was not observed in primary rat hepatocytes exposed in vitro to  $10^{-5}$ - $10^{-2}$  M DEHP or in vivo by a single gavage dose of 5 g/kg 2, 15 or 24 h prior to isolation of hepatocytes. Dietary concentration of 2% DEHP led to a marked proliferation of peroxisomes after 4 weeks. Neither this treatment nor the additional administration of a single gavage dose of 5 g/kg DEHP 15 h before sacrifice to animals fed the 2% diet for 8 weeks, or 4 weeks with or without pretreatment with 3-amino-1,2,4-triazole (to inhibit endogenous catalase activity) induced any detectable DNA-repair in hepatocytes.

### 7.3 DNA-binding

Radioactivity from carbonyl-labeled DEHP did not associate with purified protein, RNA or DNA from rat liver in vivo. Label from 2-ethyl-(1-<sup>14</sup>C)-hexyl-labeled DEHP or MEHP did appear to associate strongly with purified DNA, but label from free <sup>14</sup>C-labeled 2-ethylhexanol did not. However the apparent binding of DEHP and MEHP was not further characterized (8).

In a similar study (57) DEHP radiolabeled in different positions was administered orally to female F344 rats with or without pretreatment for 4 weeks with 1% unlabeled DEHP in the diet. Administration of <sup>14</sup>C-carboxylate-labeled DEHP resulted in no measurable DNA radioactivity while radioactivity clearly was measurable after administration of DEHP, <sup>14</sup>C- or <sup>3</sup>H-labeled in the alcohol moiety, as well as after administration of 2-ethyl(1-<sup>14</sup>C)hexanol. HPLC analysis showed that the normal nucleosides had incorporated radiolabel, while fractions expected to contain carcinogen-modified nucleoside adducts did not contain any radioactivity.

DNA isolated from the livers of male F344 rats administered 2000 mg/kg DEHP daily by gavage for 3 days was analyzed for possible carcinogen-DNA adducts by the <sup>32</sup>P-postlabeling technique (43). No adducts were detected in the DNA, which also was the case when DNA from hepatocytes exposed to 10<sup>-3</sup> M DEHP in vitro for 4 h was analyzed.

### 7.4 Chromosomal Effects

Chromosomal effects of DEHP have mainly been studied in vitro, although some studies on the induction of micronuclei in peripheral blood erythrocytes of mice have been published.

DEHP did not induce any increase in the level of sister chromatid exchange (SCE) in Chinese hamster ovary (CHO) cells, treated for 1 hour, either with or without S9, with doses up to 10 mM (32). On the other hand, MEHP has been reported to induce SCEs in V79 cells treated with 25 or 50 µg/ml for 24 hours, and 1500 µg/ml for 3 hours (115). MEHP also induced chromosomal aberrations in CHO cells and RL<sub>4</sub> cells (from rat liver), but only at cytotoxic concentrations in CHO cells (1.0 and 1.3 mM; ± S9). MEHP was less

toxic to RL<sub>4</sub> cells and nontoxic concentrations between 2.0 and 6.5 mM gave a dose-related increase in aberrations (75).

The induction of aneuploidy by DEHP was investigated both in mammalian cells and in fungi in the IPCS study (9). The mammalian assays, using either a fibroblast cell line or Chinese hamster primary liver cells, gave a positive response at 50 µg/ml and at doses between 5 and 50 µg/ml, respectively. Two out of four studies using fungi were also positive and the consensus was that DEHP had the potential of inducing aneuploidy in vitro in both fungi and mammalian cells.

DEHP and its major metabolites have also been tested for their potential of inducing micronuclei. DEHP (0.6, 3.0 or 6.0 g/kg/day/5 days) failed to induce micronuclei in peripheral blood erythrocytes of B<sub>6</sub>C<sub>3</sub>F<sub>1</sub> male mice during a 4 week period (32). Negative results were also obtained in another mouse micronucleus test after both a single dose 5 g/kg, as well as after multiple doses of 5 g/kg/day. In this study MEHP and 2-ethylhexanol were also negative (10).

#### 7.5 Cell Transformation

DEHP induced cellular transformation has been studied in several different experimental systems. In a test program (10) the BALB/3T3 cell transformation assay was used with and without rat primary hepatocyte activation. DEHP (0.875-1 ml/ml) as well as the two metabolites MEHP and 2-ethylhexanol were reported negative. On the other hand, the majority of transformation tests in the IPCS study (9) were positive for DEHP. Negative results were obtained with BALB/c-3T3 cells, while a study measuring the enhancement of viral transformation of Syrian hamster embryo (SHE) cells, was considered to be inconclusive. Positive responses were obtained by four other investigators (9), using SHE cells, with doses of 1-300 µg/ml (two different laboratories), embryonic mouse fibroblasts at 1000 µg/ml with S9 and 10 µg/ml without S9, or retrovirus-infected Fischer rat embryo cells, at the highest dose tested, 2000 µg/ml. In a separate study (115) both DEHP (7500 mg/kg and 15000 mg/kg) and MEHP (375 mg/kg and 750 mg/kg) induced morphological transformation, as well as chromosomal aberrations, in SHE cells after transplacental administration.

The induction of anchorage independence in JB6 mouse epidermal cells has been used to study tumour promoting activity in vitro. This assay gave a positive response for DEHP at concentrations of 500 to 20,000 ppm/ml culture medium (31). In another study (120) with the same model system, both DEHP ( $1.3\text{--}51 \times 10^{-6}$  M) and MEHP ( $2\text{--}5 \times 10^{-8}$  M) were positive, while 2-ethylhexanol ( $4\text{--}77 \times 10^{-7}$  M) was negative.

There are also a few studies on DEHP induced inhibition of metabolic cooperation, which may be indicative of the promoting potential of a substance. Metabolic cooperation in Chinese hamster V79 cells was not inhibited by DEHP at non-cytotoxic concentrations, 300 nM (0.12 µg/ml) or below (48). In the IPCS study (9) one investigator reported an inhibition in V79 cells at non-cytotoxic concentrations of DEHP (25-200 µg/ml in two separate experiments and 5-25 µg/ml in another), while another investigator, using V79 cells in a microassay method, detected a slight but non-significant increased inhibition with doses between 10 and 200 nM.

#### 7.6 Summary

Numerous studies on possible genotoxic effects of DEHP have been performed using different short term tests. The results indicate that DEHP is not a direct acting genotoxic agent in vitro. With few exceptions mutagenicity studies both in bacteria and eucaryotic cells have been negative and essentially no positive response has been obtained in assays measuring DNA-damage. This is also consistent with results obtained in DNA-binding studies, which have been unable to detect any covalent binding of DEHP or its metabolites to DNA. Tests measuring chromosomal aberrations and SCE have also largely been negative. However, MEHP has been found to be a clastogen in vitro but it has been argued that this effect could be related to the surface active properties of MEHP. A few in vivo assays, measuring the induction of micronuclei, UDS or strand breaks in DNA, have also been performed with negative results. Neither could any mutagenic metabolites be detected in the urine from DEHP treated rats, using a Salmonella assay. In contradiction to these negative results, a few in vitro assays measuring aneuploidy have mainly been positive. Different cell transformation assays that have been performed have also mainly been positive. Although DEHP is not as thoroughly tested in these assays, the mainly positive results obtained indicate that DEHP can affect the cellular genome in vitro.

## 8 CARCINOGENICITY

In a carcinogenicity study (47, 62) groups of 50 male and female Fischer 344 rats and male and female B6C3F<sub>1</sub> mice were fed diets containing 6000 or 12000 (rats) or 3000 or 6000 (mice) mg DEHP/kg for 103 consecutive weeks. Concurrent controls (50 of each sex and species) were fed diet without the addition of DEHP. All of the animals were given control diet for 1-2 weeks after 103 weeks of treatment and were then killed and examined both grossly and microscopically. Food and water were supplied ad libitum. The administered concentrations of DEHP were estimated to be half maximally tolerated doses and maximally tolerated doses.

Table I. Carcinogenic effect of DEHP on the liver (from ref 62)

	Control	Low dose	High dose
<b>Hepatocellular carcinoma</b>			
Male rats	1/50	1/49	5/49
Female rats	0/50	2/49	8/50
Male mice	9/50	14/48	19/50
Female mice	0/50	7/50	17/50
<b>Neoplastic nodules</b>			
Male rats	2/50	5/49	7/49
Female rats	0/50	4/49	5/50
<b>Hepatocellular adenoma</b>			
Male mice	6/50	11/48	10/50
Female mice	1/50	5/50	1/50



Under these conditions, DEHP caused an increased incidence in female rats and male and female mice of hepatocellular carcinomas, and an increased incidence in male rats of either hepatocellular carcinomas or neoplastic nodules. (See Table I.) Twenty of the 57 hepatocellular carcinomas in the DEHP-treated mice (sexes and doses combined) had metastasized to the lung. The 9 hepatocellular carcinomas in control male mice are said to be within a normal range (62).

The reported decreased incidence of tumors of the thyroid, pituitary, and testis could be related to an increased endocrine activity of the pituitary gland (62).

The carcinogenicity of DEHP was supported by the results of another chronic study (78) even though the group sizes were small in that study.

IARC (45) has made the evaluation that there is sufficient evidence for the carcinogenicity of DEHP phthalate in mice and rats, based on a significantly increased incidence of liver cell tumors in animals of both species, and an observed dose-response relationship.

Two other long-term studies have been performed (21, 44), but due to the small numbers of animals used, the studies are inadequate to assess the carcinogenic potential of DEHP.

## 9 SPECIAL STUDIES

Since DEHP in most test systems lacks genotoxic activity it has been hypothesized that the carcinogenic effect is exerted during the promotion phase of hepatocarcinogenicity. DEHP has therefore been tested in several initiation/promotion experiments in rats and mice where the end point has been the number and/or volume of foci of altered liver cells. As expected, DEHP lacks initiating activity in these experiments (78, 120). DEHP is a probable promoter of such foci in the mouse liver and accordingly a possible tumor promoter in the mouse (120). In the rat, however, DEHP does not promote altered foci during two years of feeding (78) and even seems to accelerate the regression or inhibit the appearance of some kinds of such

foci (28, 29). The reasons for this, as well as for the divergent responses in rats and mice, are unclear.

Extrapolation from mouse data to humans is considered impossible since the mechanism of altered foci promotion in mice is unknown.

## 10 EFFECTS ON MAN

Two adult subjects who were given 5 or 10 g of DEHP experienced no untoward effects apart from mild gastric disturbances and moderate catharsis at the 10 g dose (95). Three cases of non-specific hepatitis were described among 27 hemodialysis patients with terminal renal failure. The PVC blood tubings used released DEHP in an amount of 10-20 mg/l perfusate. The symptoms and signs of hepatitis disappeared rapidly when the use of tubings not containing DEHP was resumed (63).

There are very few data on effects on man of specific DEHP occupational exposure on man.

Two studies report symptoms and signs of polyneuropathy among 47 out of 147 and 12 out of 23 workers examined in a Russian PVC-processing plant and an Italian plant for phthalate production. The workers were exposed to mixed phthalates and DEHP was a minor exposure, at least in the Russian plant. The total phthalate air concentrations recorded varied between 1.7 and 66 mg/m<sup>3</sup> and 1 and 60 mg/m<sup>3</sup> respectively (36, 59).

In a study from a Swedish PVC-processing factory the presence of symptoms and signs from the peripheral nervous system were examined among 54 male workers exposed mainly to DEHP, diisodecylphthalate and some butylbenzylphthalate. The workers were divided into three groups of equal size with mean phthalate exposures 0.1, 0.2 and 0.7 mg/m<sup>3</sup>, respectively. None of the outcome variables studied showed a significant association with exposure (64). No subject reported work related obstructive lung disease and conventional lung function tests showed no association with exposure.

Several biochemical parameters showed significant associations with exposure. There was a slight decrease of the hemoglobin level with time of

employment as well as exposure in the last year. Alpha-1-antitrypsin in serum increased slightly with time of employment and immunoglobulin A in serum rose with rising exposure during the last year (64).

One case of occupational asthma due to DEHP has been reported in a PVC-processing worker (17). When the patient was exposed to DEHP in an inhalation chamber an asthmatic reaction was elicited. The action was inhibited by prior administration of sodium chromoglycate.

A study of blood lipids, serum activities of liver enzymes and routine hematological tests was carried out among workers in a German plant for DEHP production. The study was negative and uninformative due to lack of a control group and very low exposures, below  $0.16 \text{ mg/m}^3$  (112).

Thiess & Flieg (110) investigated the frequency of chromosome aberrations in 10 workers engaged in DEHP production in the same plant from 10-30 years. There was no increase in chromosome aberrations compared to control groups but air exposure levels were very low,  $0.09\text{-}0.16 \text{ mg/m}^3$ .

A mortality study of 221 workers exposed to DEHP in the same plant was also conducted in which there were only 8 deaths. The study is uninformative due to small cohort size, short follow-up and low exposures (111).

IARC (45) concluded that no adequate data were available to assess the carcinogenicity of DEHP to humans.

## 11 CONCLUSIONS

Data on occupational exposure concentrations are limited. Concentrations of DEHP up to about  $5 \text{ mg/m}^3$  have been reported but usually the industrial levels of DEHP are below  $1 \text{ mg/m}^3$ . In some cases the total phthalate exposure concentrations have been recorded to be between 0.1 and  $60 \text{ mg/m}^3$ .

Dose dependent kinetics of DEHP or its metabolites have been indicated in several studies as well as marked animal species differences in metabolism. Induction phenomena have also been described. These facts complicate

extrapolation from animal studies to man and also extrapolations from high doses to low doses.

Data on effects of DEHP on humans are scarce and do not permit any conclusions on dose-effect or dose-response relationships.

There are no data on health effects from DEHP exposure outside the occupational environment. A few studies of workers exposed to phthalate mixtures have been published. However, no consistent findings are reported.

One case-report suggests that DEHP could cause asthma.

In animals few inhalation studies have been performed. The oral and intraperitoneal LD<sub>50</sub> values exceed 25 g/kg, which indicate that DEHP has low acute toxicity. Prolonged dosing of DEHP produces hepatomegaly and proliferation of peroxisomes. In the rat the no-effect level for feeding studies appears to be about 0.01% in the diet. The no-effect level in the rat for testicular atrophy fed DEHP is about 0.3-0.5% in the diet, and is more pronounced in young animals. Hamsters and mice appears to be more resistant to the testicular effects of DEHP.

In rats 2% DEHP in the diet throughout gestation produced an increased incidence of resorptions but not of malformations. In the mouse, however, 0.1% throughout pregnancy increased the incidence of embryoletality and abnormalities. Days 7-9 of gestation were most sensitive. The dose, 0.05 g/kg, to mice which induced fetotoxicity, is not expected to induce maternal toxicity. Contradictory results are obtained in fertility studies with DEHP.

Results from several different genotoxicity tests indicate that DEHP and its major metabolites do not exhibit any direct genotoxic effect in either bacteria, fungi or mammalian cells in vitro. This has also been confirmed in binding studies with DEHP, which indicated that DEHP and its metabolites do not interact covalently with DNA. However, it has been established that DEHP has the potential of inducing aneuploidi in fungi, as well as in mammalian cells in vitro. These results, together with several positive results obtained in different cell transformation assays, indicate that DEHP can effect the cellular genome in vitro.

DEHP acts as a tumor promoter in mouse liver but not in rat liver in established in vivo test systems.

High doses (12000 mg/kg in rats; 6000 mg/kg in mice) of DEHP in a feeding study resulted in increased incidences of hepatocellular carcinoma.

Based on animal data it can be concluded that DEHP is carcinogenic and teratogenic. Due to a lack of human data the degree of risk to humans can not be evaluated, but DEHP should be considered as potentially carcinogenic and teratogenic to humans.

## 12 SUMMARY

### 12.1 Summary in English

P. Garberg, J. Högberg, I. Lundberg, P. Lundberg. NIOH and NIOSH basis for an occupational health standard: Di(2-ethylhexyl)phthalate. Arbete och Hälsa 1989:25, pp 1-53.

The document is a survey of the literature with regard to health effects of exposure to DEHP. Data on occupational exposure levels and effects on humans are limited. DEHP is hydrolyzed to MEHP and further oxidized. No accumulation has been observed.

DEHP has a low acute toxicity. Long-term exposure to rats causes testicular atrophy. DEHP is carcinogenic and teratogenic in rodents. Most genotoxicity tests have been negative. However, certain tests have indicated that DEHP may affect the genome in vitro.

Key words: DEHP, testicular atrophy, teratogenicity, carcinogenicity, genotoxicity

### 12.2 Sammanfattning på svenska

P. Garberg, J. Högberg, I. Lundberg, P. Lundberg. NIOH and NIOSH basis for an occupational health standard: Di(2-ethylhexyl)phthalate. Arbete och Hälsa 1989:25, s 1-53.

Dokumentet utgör en sammanställning av litteraturuppgifter beträffande hälsoeffekter vid exponering för DEHP. Data på exponeringsnivåer i arbetsmiljö och effekter på människa är begränsade. DEHP hydrolyseras till MEHP som oxideras vidare. Någon ackumulering har inte observerats.

DEHP har låg akuttoxicitet. Långtidsexponering av råtta orsakar testikelatrofi. DEHP är carcinogen och teratogen för gnagare. De flesta genotoxicitetstest har givit negativa resultat. Vissa test antyder att DEHP dock kan påverka genomet in vitro.

Nyckelord: DEHP, testikelatrofi, teratogenicitet, carcinogenicitet, genotoxicitet

### 13 REFERENCES

1. Agarwal DK. Letters to the editor, Toxicol Appl Pharmacol 82 (1986) 383-387.
2. Agarwal DK, Agarwal S, Seth PK. Effect of di-(2-ethylhexyl) phthalate on drug metabolism, lipid peroxidation, and sulfhydryl content of rat liver. Drug Metabol Disp 10 (1982) 77-80.
3. Agarwal DK, Lawrence WH, Autian J. Antifertility and mutagenic effects in mice from parenteral administration of di-2-ethylhexyl phthalate (DEHP). J Toxicol Environ Health 16 (1985) 71-84.
4. Agarwal DK, Lawrence WH, Nunez LJ, Autian J. Mutagenicity evaluation of phthalic acid esters and metabolites in Salmonella typhimurium cultures. J Toxicol Environ Health 16 (1985) 61-69.
5. Albro PW. The biochemical toxicology of di-(2-ethylhexyl) and related phthalates: testicular atrophy and hepatocarcinogenesis. In Hodgson E, Bend JR and Philpot RM (Eds), Reviews in Biochemical Toxicology, 8, pp 73-119. Elsevier, New York 1987.
6. Albro PW, Thomas RO. Enzymatic hydrolysis of di(2-ethylhexyl)phthalate by lipases. Biochim Biophys Acta 360 (1973) 380-390.
7. Albro PW, Thomas R, Fishbein L. Metabolism of diethylhexyl phthalate by rats. Isolation and characterization of the urinary metabolites. J Chromatography 76 (1973) 321-330.
8. Albro PW, Corbett JT, Schroeder JL, Jordan S, Matthews HB. Pharmacokinetics, interactions with macromolecules and species differences in metabolism of DEHP. Environ Health Perspect 45 (1982) 19-25.
9. Asbhy J, De Serres FJ, Draper M, Ishidate Jr M, Margolin BH, Matter BE, Shelby MD (Eds). Progress in Mutation Research. Evaluation of short-term tests for carcinogens. Report of the International

Programme on Chemical Safety's Collaborative Study on In Vitro Assays, Vol 5, 752 pp. Elsevier Science Publishers, Amsterdam, New York 1985.

10. Astill B, Barber E, Lington A, Moran E, Mulholland A, Robinson E, Scheider B. Chemical industry voluntary test program for phthalate esters: Health effect studies. *Environ Health Perspect* 65 (1986) 329-336.
11. Bally MB, Opheim DJ, Shertzer HG. Di-(2-ethylhexyl) phthalate enhances the release of lysosomal enzymes from alveolar macrophages during phagocytosis. *Toxicology* 18 (1980) 49-60.
12. Barney GH, Orgebin-Crist MC, Macapinlac MP. Genesis of esophageal parakeratosis and histologic changes in the testes of the zinc-deficient rat and their reversal by zinc repletion. *J Nutr* 95 (1969) 526-534.
13. Bell FP, Patt CS, Brundage B, Gillies PJ, Philips WA. Studies on lipid biosynthesis and cholesterol content of liver and serum lipoproteins in rats fed various phthalate esters. *Lipids* 3 (1978) 66-74.
14. Bell FP, Patt CS, Gillies PJ. Effect of phthalate esters on serum cholesterol and lipid biosynthesis in liver, testes, and epididymal fat in the rat and rabbit. *Lipids* 3 (1978) 673-678.
15. Bell FP, Wang S, Mendoza AR, Nishizawa EE. Platelet function and platelet lipid biosynthesis in rats and rabbits fed the plasticiser DEHP. *Bull Environ Contam Toxicol* 23 (1979) 306-310.
16. Bridges JW. Frontiers in biochemical toxicology. *TIPS*, 6 (suppl) (1985) 11-15.
17. Brunetti G, Moscato G. Asma bronchiale da esposizione professionale a dioctilftalato. *Med Lavoro* 75 (1984) 120-124.



18. Butterworth BE, Bermudez E, Smith-Oliver T, Earle L, Cattley R, Martin J, Popp JA, Strom S, Jirtle R, Michalopoulos G. Lack of genotoxic activity of di(2-ethylhexyl) phthalate (DEHP) in rat and human hepatocytes. *Carcinogenesis* 5 (1984) 1329-1335.
19. Butterworth BE, Loury DJ, Smith-Oliver T, Cattley RC. The potential role of chemically induced hyperplasia in the carcinogenic activity of the hypolipidemic carcinogens. *Toxicol Ind Health* 3 (1987) 129-148.
20. Calley D, Autian J, Guess WL. Toxicology of a series of phthalate esters. *J Pharm Sci* 55 (1966) 158-162.
21. Carpenter CP, Weil CS, Smyth HF. Chronic oral toxicity of di(2-ethylhexyl) phthalate for rats, guinea pigs, and dogs. *Arch Ind Hyg* 8 (1953) 219-226.
22. Chu I, Secours VE, Marino IA, Villeneuve DC, Valli VE. Sub-acute and sub-chronic toxicity of mono-2-ethylhexyl phthalate in the rat. *Arch Environ Contam Toxicol* 10 (1981) 271-280.
23. Clayton GD, Clayton FE. *Patty's Industrial Hygiene and Toxicology*, 3rd ed., Wiley-Interscience, New York 1981.
24. Curstedt T, Sjövall J. Individual molecular species of phosphatidylcholines and phosphatidylinositols in livers of rats fed bis(2-ethylhexyl) phthalate. *Med Biol* 61 (1983) 219-222.
25. Curto KA, McCafferty RE, Donovan MP, Thomas JA. Further studies on the effects of the phthalate acid esters (PAE) on rat male reproductive organs. *Toxicologist* 2 (1982) 71.
26. Curto KA, Thomas JA. Comparative effects of diethylhexyl phthalate or monoethylhexyl phthalate on male mouse and rat reproductive organs. *Toxicol Appl Pharmacol* 62 (1982) 121-125.

27. Daniel JW, Bratt H. The absorption, metabolism and tissue distribution of di(2-ethylhexyl) phthalate in rats. *Toxicology* 2 (1974) 51-65.
28. Deangelo AB, Garrett CT. Inhibition of development of preneoplastic lesions in the livers of rats fed a weakly carcinogenic environmental contaminant. *Cancer Lett* 20 (1983) 199-205.
29. Deangelo AB, Garrett CT, Queral AE. Inhibition of phenobarbital and dietary choline deficiency promoted preneoplastic lesions in rat liver by environmental contaminant di(2-ethylhexyl) phthalate. *Cancer Lett* 23 (1984) 323-330.
30. Divincenzo GD, Hamilton ML, Mueller KR, Donish WH, Barber ED. Bacterial mutagenicity testing of urine from rats dosed with 2-ethylhexanol derived plasticizers. *Toxicology* 34 (1985) 247-259.
31. Diwan BA, Ward JM, Rice JM, Colburn NH, Spangler EF. Tumor-promoting effects of di(2-ethylhexyl) phthalate in JB6 mouse epidermal cells and mouse skin. *Carcinogenesis* 6 (1985) 343-347.
32. Douglas RG, Hugenholtz AP, Blakey DH. Genetic toxicology of phthalate esters: mutagenic and other genotoxic effects. *Environ Health Perspect* 65 (1986) 255-262.
33. Fishbein L, Albro PW. Chromatographic and biological aspects of the phthalate esters. *J Chromatogr* 70 (1972) 365-412.
34. Foster PMD, Lake BG, Cook MW, Thomas LV, Gangolli SD. Structure-activity requirements for the induction of testicular atrophy by butyl phthalates in immature rats: effects on testicular zinc content. *Adv Exp Med Biol* 136A (1982) 445-452.
35. Foster PMD, Thomas LV, Cook MW, Walters DG. Effect of di-n-pentyl phthalate treatment on testicular steroidogenic enzymes and cytochrome P-450 in the rat. *Toxicol Lett* 15 (1983) 265-271.

36. Gilioli R, Bulgheroni C, Terrana T, Filippini G, Massetto N, Boeri R. Studio neurologico trasversale e longitudinale di una popolazione operaia addetta alla produzione di ftalati. Med Lavoro 69 (1978) 620-631.
37. Gollamudi R, Lawrence WH, Rao RH, Autian J. Effects of phthalic acid esters on drug metabolizing enzymes of rat liver. J Appl Toxicol 5 (1985) 368-371.
38. Gray TJB, Beamand, JA. Effect of some phthalate esters and other testicular toxins on primary cultures of testicular cells. Fd Chem Toxicol 22 (1984) 123-131.
39. Gray TJB, Butterworth KR. Testicular atrophy produced by phthalate esters. Arch Toxicol Suppl 4 (1980) 452-455.
40. Gray TJB, Butterworth KR, Gaunt IF, Grasso P, Gangolli SD. Short-term toxicity study of di-(2-ethylhexyl) phthalate in rats. Fd Cosmet Toxicol 15 (1977) 389-399.
41. Gray JB, Gangolli SD. Aspects of the testicular toxicity of phthalate esters. Environ Health Perspect 65 (1986) 229-235.
42. Gray TJB, Rowland IR, Foster PMD, Gangolli SD. Species differences in the testicular toxicity of phthalate esters. Toxicol Lett 11 (1982) 141-147.
43. Gupta RC, Goel SK, Earley K, Singh B, Reddy JK. <sup>32</sup>P-Postlabeling analysis of peroxisome proliferator-DNA adduct formation in rat liver in vivo and hepatocytes in vitro. Carcinogenesis 6 (1985) 933-936.
44. Harris RS, Hodge HC, Maynard EA, Blanchet Jr HJ. Chronic oral toxicity of 2-ethylhexyl phthalate in rats and dogs. Am Med Assoc Arch Ind Health 13 (1956) 259-264.
45. IARC. Monographs on the evaluation of the carcinogenic risk of chemicals to humans, Vol. 29, pp 269-294 (Some industrial chemicals and dyestuffs). IARC, Lyon, France 1982.

46. Kirby PE, Pizzarello RF, Lawlor TE, Haworth SE, Hodgson JR. Evaluation of di(2-ethylhexyl) phthalate and its major metabolites in the Ames test and L5178Y mouse lymphoma mutagenicity assay. *Environ Mutagen* 5 (1983) 657-663.
47. Kluwe W, Haseman JK, Douglas JF, Huff JE. The carcinogenicity of dietary di(2-ethylhexyl) phthalate (DEHP) in Fisher 344 rats and B6C3F<sub>1</sub> mice. *J Toxicol Environ Health* 10 (1982) 797-815.
48. Kornbrust DJ, Barfknecht TR, Ingram P, Shelburne JD. Effect of di(2-ethylhexyl)phthalate on DNA repair and lipid peroxidation in rat hepatocytes and on metabolic cooperation in Chinese hamster V-79 cells. *J Toxicol Environ Health* 13 (1984) 99-116.
49. Krauskopf LG. Studies on the toxicity of phthalates via ingestion. *Environ Health Perspect* 3 (1973) 61-72.
50. Lake BG, Gray TJB, Foster JR, Stubberfield CR, Gangolli SD. Comparative studies on di-(2-ethylhexyl) phthalate-induced hepatic peroxisome proliferation in the rat and hamster. *Toxicol Appl Pharmacol* 72 (1984) 46-60.
51. Lawrence WH, Malik M, Turner JE, Singh AR, Autian J. A toxicological investigation of some acute, short-term, and chronic effects of administering di-2-ethylhexyl phthalate (DEHP) and other phthalate esters. *Environ Res* 9 (1975) 1-11.
52. Lewandowski M, Fernandes J, Chen TS. Assessment of the teratogenic potential of plasma-soluble extracts of diethylhexyl phthalate plasticized polyvinyl chloride plastics in rats. *Toxicol Appl Pharmacol* 54 (1980) 141-147.
53. Leyder F, Boulanger P. Ultraviolet absorption, aqueous solubility, and octanol-water partition for several phthalates. *Bull Environ Contam Toxicol* 30 (1983) 152-157.

54. Lhuguenot JC, Elcombe CR. Variation selon l'espèce des phénomènes de conjugation lors du métabolisme mono-(ethyl-2-hexyl) phthalate (MEHP). *Sci Aliments* 4 (1984) 227-232.
55. Lhuguenot J-C, Mitchell AM, Milner G, Lock EA, Elcombe CR. The metabolism of di(2-ethylhexyl)phthalate (DEHP) and mono-(2-ethylhexyl) phthalate (MEHP) in rats: In vivo and in vitro dose and time dependency of metabolism. *Toxicol Appl Pharmacol* 80 (1985) 11-22.
56. Liss GM, Albro PW, Hartle RW, Stringer WT. Urine phthalate determinations as an index of occupational exposure to phthalic anhydride and di(2-ethylhexyl) phthalate. *Scand J Work Environ Health* 11 (1985) 381-387.
57. Lutz WK. Investigation of the potential for binding of di(2-ethylhexyl) phthalate (DEHP) to rat liver DNA in vivo. *Environ Health Perspect* 65 (1986) 267-269.
58. Melnick RL, Morrissey RE, Tomaszewski KE. Studies by the National Toxicology Program on di(2-ethylhexyl) phthalate. *Toxicol Ind Health* 3 (1987) 99-118.
59. Milkov LE, Aldyreva MV, Popova TB, Lopukhova KA, Makarenko YL, Malyar LM, Shakhova TK. Health status of workers exposed to phthalate plasticizers in the manufacture of artificial leather and films based on PVC resins. *Environ Health Perspect* 3 (1973) 175-178.
60. Mitchell AM, Lhuguenot J-C, Bridges JW, Elcombe CR. Identification of the proximate peroxisome proliferator(s) derived from di(2-ethylhexyl) phthalate. *Toxicol Appl Pharmacol* 80 (1985) 23-32.
61. Mitchell FE, Price SC, Hinton RH, Grasso P, Bridges JW. Time and dose-response study of the effects on rats of the plasticizer di(2-ethylhexyl) phthalate. *Toxicol Appl Pharmacol* 81 (1985) 371-392.

62. National Toxicology Program. Technical report on the carcinogenicity bioassay of di(2-ethylhexyl) phthalate. DHHS Publ No (NIH) 82-1773. Research Triangle Park, N.C. 1982.
63. Neergaard J, Nielsen B, Faubry V, Christensen DH, Nielsen OF. Plasticizers in P.V.C. and the occurrence of hepatitis in a haemodialysis unit. Scand J Urol Nephrol 5 (1971) 141-145.
64. Nielsen J, Åkesson B, Skerfving S. Phthalate ester exposure - air levels and health of workers processing polyvinylchloride. Am Ind Hyg Assoc J 46 (1985) 643-647.
65. Nikonorow M, Mazur H, Piekacz H. Effect of orally administered plasticizers and polyvinyl chloride stabilizers in the rat. Toxicol Appl Pharmacol 26 (1973) 253-259.
66. NIOSH. Registry of Toxic Effects of Chemical Substances, 1983-84 Supplement, Vol. II, p 1442. National Institute for Occupational Safety and Health, Cincinnati, Ohio 1985.
67. Oishi S. Testicular atrophy of rats induced by di-2-ethylhexyl phthalate: effects of vitamin A and zinc concentrations in the testis, liver and serum. Toxicol Lett 20 (1984) 75-78.
68. Oishi S. Reversibility of testicular atrophy induced by di(2-ethylhexyl) phthalate in rats. Environ Res 36 (1984) 160-169.
69. Oishi S, Hiraga K. Effect of phthalic acid esters on gonadal function in male rats. Bull Environ Contam Toxicol 21 (1979) 65-67.
70. Oishi S, Hiraga K. Testicular atrophy induced by phthalic esters: effect on testosterone and zinc concentrations. Toxicol Appl Pharmacol 53 (1980) 35-41.
71. Oishi S, Hiraga K. Effect of phthalic acid monoesters on mouse testes. Toxicol Lett 6 (1980) 239-242.

72. Oishi S, Hiraga K. Effects of monoesters of o-phthalic acid on serum lipid composition in rats. *Toxicol Lett* 14 (1982) 79-84.
73. Oishi S, Hiraga K. Testicular atrophy induced by di-2-ethylhexyl phthalate: effect of zinc supplement. *Toxicol Appl Pharmacol* 70 (1983) 43-48.
74. Okyama T. Effects of phthalate esters on glucose-6-phosphate dehydrogenase and other enzymes in vitro. *Toxicol Appl Pharmacol* 40 (1977) 355-364.
75. Phillips BJ, Anderson D, Gangolli SD. Studies on the genetic effects of phthalic acid esters on cells in culture. *Environ Health Perspect* 65 (1986) 263-266.
76. Pollack GM, Buchanan JF, Slaughter RL, Kohli RK, Shen DD. Circulating concentrations of di(2-ethylhexyl) phthalate and its de-esterified phthalic acid products following plasticizer exposure in patients receiving hemodialysis. *Toxicol Appl Pharmacol* 79 (1985) 257-267.
77. Pollack GM, Li RCK, Ermer JC, Shen DD. Effects of route of administration and repetitive dosing on the disposition kinetics of di(2-ethylhexyl) phthalate and its mono-de-esterified metabolite in rats. *Toxicol Appl Pharmacol* 79 (1985) 246-256.
78. Popp JA, Garvey LK, Cattley RC. In vivo studies on the mechanism of di(2-ethylhexyl) phthalate carcinogenesis. *Toxicol Ind Health* 3 (1987) 151-163.
79. Quackenbos HM. Plasticisers in vinyl chloride resins. Migration of plasticiser. *Ind Eng Chem* 46 (1953) 1335-1344.
80. Reddy JK, Lalwani ND. Carcinogenesis by hepatic peroxisome proliferators: evaluation of the risk of hypolipidemic drugs and industrial plasticizers to humans. *CRC Crit Rev Toxicol* 12 (1983) 1-58.

81. Reddy JK, Reddy MK, Usman MI, Lalwani ND, Rao MS. Comparison of hepatic peroxisome proliferative effect and its implication for hepatocarcinogenicity of phthalate esters, di(2-ethylhexyl) phthalate, and di(2-ethylhexyl) adipate with a hypolipidemic drug. *Environ Health Perspect* 65 (1986) 317-327.
82. Reddy JK, Warren JR, Reddy MK, Lalwani ND. Hepatic and renal effects of peroxisome proliferators: biological implications. *Ann N.Y. Acad Sci* 386 (1982) 81-110.
83. Rhodes C, Orton TC, Pratt IS, Batten PL, Bratt H, Jackson SJ, Elcombe CR. Comparative pharmacokinetics and subacute toxicity of di(2-ethylhexyl)phthalate (DEHP) in rats and marmosets: extrapolation of effects in rodents to man. *Environ Health Perspect* 65 (1986) 299-308.
84. Rubin RJ, Chang JFC. Effect of the intravenous administration of the solubilized plasticizer di(2-ethylhexyl) phthalate on the lung and on survival of transfused rats. *Toxicol Appl Pharmacol* 45 (1978) 230.
85. Rubin RJ, Jaeger RJ. Some pharmacologic and toxicologic effects of di-2-ethylhexyl phthalate (DEHP) and other plasticizers. *Environ Health Perspect* 3 (1973) 53-59.
86. Ruddick JA, Villeneuve DC, Chu I, Nestmann E, Miles D. An assessment of the teratogenicity in the rat and mutagenicity in *Salmonella* of mono-2-ethylhexyl phthalate. *Bull Environ Contam Toxicol* 27 (1981) 181-186.
87. Sakurai T, Miyazawa S, Hashimoto T. Effect of di(2-ethylhexyl) phthalate administration on carbohydrate and fatty acid metabolism in rat liver. *J Biochem* 83 (1978) 313-320.
88. Sax NI. *Dangerous Properties of Industrial Materials*, 6th ed., Van Nostrand Reinhold Co, New York 1984.
89. Saxena DK, Srivastava SP, Chandra SV, Seth PK. Testicular effects of di(2-ethylhexyl) phthalate (DEHP): histochemical and histopathological alterations. *Ind Health* 23 (1985) 191-198.



90. Schmid P, Schlatter C. Excretion and metabolism of di(2-ethylhexyl) phthalate in man. *Xenobiotica* 15 (1985) 251-256.
91. Schmidt JG, Garvin PJ, Leestma JE. Effect of vehicle on the response to intravenous di-(2-ethylhexyl) phthalate (DEHP) in rats. *Toxicol Appl Pharmacol* 33 (1975) 169.
92. Schulz CO, Rubin RJ. Distribution, metabolism, and excretion of di-2-ethylhexyl phthalate in the rat. *Environ Health Perspect* 3 (1973) 123-129.
93. Schulz CO, Rubin RJ, Hutchins GM. Acute lung toxicity and sudden death in rats following the intravenous administration of the plasticizer, di(2-ethylhexyl)-phthalate, solubilized with Tween surfactants. *Toxicol Appl Pharmacol* 33 (1975) 514-525.
94. Seth PK, Srivastava SP, Agarwal DK, Chandra SV. Effect of di-2-ethylhexyl phthalate (DEHP) on rat gonads. *Environ Res* 12 (1976) 131-138.
95. Shaffer CB, Carpenter CP, Smyth Jr HF. Acute and subacute toxicity of di(2-ethylhexyl) phthalate with note upon its metabolism. *J Ind Hyg Toxicol* 27 (1945) 130-135.
96. Shindo Y, Osumi T, Hashimoto T. Effects of administration of di(2-ethylhexyl) phthalate on rat liver mitochondria. *Biochem Pharmacol* 27 (1978) 2683-2688.
97. Shiota K, Chou MJ, Nishimura H. Embryotoxic effects of di-2-ethylhexyl phthalate (DEHP) and di-n-butyl phthalate (DBP) in mice. *Environ Res* 22 (1980) 245-253.
98. Shiota K, Mima S. Assessment of the teratogenicity of di(2-ethylhexyl) phthalate and mono(2-ethylhexyl) phthalate in mice. *Arch Toxicol* 56 (1985) 263-266.

99. Shiota K, Nishimura H. Teratogenicity of di(2-ethylhexyl) phthalate (DEHP) and di-n-butyl phthalate (DBP) in mice. *Environ Health Perspect* 45 (1982) 65-70.
100. Shukla RR, Albro PW. In vitro modulation of protein kinase c activity by environmental chemical pollutants. *Biochem Biophys Res Commun* 142 (1987) 567-572.
101. Singh AR, Lawrence WH, Autian J. Teratogenicity of phthalate esters in rats. *J Pharm Sci* 61 (1972) 51-55.
102. Sittig M. *Handbook of Toxic and Hazardous Chemicals and Carcinogens*, 2nd ed, Noyes Publications, Park Ridge, New Jersey 1985.
103. Sjöberg P, Bondesson U, Kjellen L, Lindquist NG, Montin G, Plöen L. Kinetics of di-(2-ethylhexyl) phthalate in immature and mature rats and effect on testis. *Acta Pharmacol Toxicol* 56 (1985) 30-37.
104. Sjöberg P, Bondesson U, Sedin G, Gustafsson J. Dispositions of di- and mono-(2-ethylhexyl) phthalate in newborn infants subjected to exchange transfusions. *Eur J Clin Invest* 15 (1985) 430-436.
105. Sjöberg P, Lindquist NG, Montin G, Plöen L. Effect of repeated intravenous infusions of the plasticizer di-(2-ethylhexyl) phthalate in young male rats. *Arch Toxicol* 58 (1985) 78-83.
106. Sjöberg P, Lindquist NG, Plöen L. Age-dependent response of the rat testes to di(2-ethylhexyl) phthalate. *Environ Health Perspect* 65 (1986) 234-242.
107. Sjöberg P, Plöen L, Bondesson U, Lindquist NG. Testicular toxicity of di-(2-ethylhexyl) phthalate in rats of different ages. *Teratology* 26 (1982) 20A.
108. Swenerton H, Hurley LS. Teratogenic effects of a chelating agent and their prevention by zinc. *Science* 173 (1971) 62-63.

109. Teirlynck OA, Belpaire F. Disposition of orally administered di-(2-ethylhexyl) phthalate and mono-(2-ethylhexyl) phthalate in the rat. Arch Toxicol 57 (1985) 226-230.
110. Thiess AM, Flieg I. Chromosomenuntersuchungen bei Mitarbeitern mit Exposition gegenüber Di-2-äthylhexylphthalat (DOP). Zbl Arbeitsmed 28 (1978) 351-355.
111. Thiess AM, Frentzel-Beyme R, Wieland R. [Mortality study in workers exposed to di-2-ethylhexyl phthalate (DOP).] (Ger) In: Möglichkeiten und Grenzen des Biological Monitoring. Arbeitsmedizinische Probleme des Dienstleistungsgewerbes. Arbeitsmedizinische Kolloquium, Frankfurt, May 1978, pp 155-164. A.W. Gertner, Stuttgart 1978.
112. Thiess AM, Korte A, Flieg H. [Morbidity studies on workers exposed to di(2-ethylhexyl)phthalate (DEHP).] (Ger) In: Möglichkeiten und Grenzen des Biological Monitoring. Arbeitsmedizinische Probleme des Dienstleistungsgewerbes. Arbeitsmedizinische Kolloquium, Frankfurt, May 1978, pp 137-154. A.W. Gertner, Stuttgart 1978.
113. Thomas JA, Darby TD, Wallin RF, Garvin PJ, Martis L. A review of the biological effects of di-(2-ethylhexyl) phthalate. Toxicol Appl Pharmacol 45 (1978) 1-27.
114. Thomas JA, Schein LG, Gupta PK, Mc Cafferty RE, Felice PR, Donovan MP. Failure of monoethylhexyl phthalate to cause teratogenic effects in offspring of rabbits. Toxicol Appl Pharmacol 51 (1979) 523-528.
115. Tomita I, Nakamura Y, Aoki N, Inui N. Mutagenic/carcinogenic potential of DEHP and MEHP. Environ Health Perspect 45 (1982) 119-125.
116. Tomita I, Nakamura Y, Yagi Y, Tutikawa K. Teratogenicity/fetotoxicity of DEHP in mice. Environ Health Perspect 45 (1982) 71-75.

117. Tomita I, Nakamura Y, Yagi Y, Tutikawa K. Fetotoxic effects of mono-2-ethylhexyl phthalate (MEHP) in mice. *Environ Health Perspect* 65 (1986) 249-254.
118. Villeneuve DC, Franklin CA, Chu I, Yagminas A, Marino IA, Ritter L, Ruddick JA. Toxicity studies on mono-2-ethylhexyl phthalate. *Toxicol Appl Pharmacol* 45 (1978) 250-251.
119. Walseth F, Toftgård R, Nilsen OG. Phthalate esters I: Effects on cytochrome P-450 mediated metabolism in rat liver and lung, serum enzymatic activities and serum protein levels. *Arch Toxicol* 50 (1982) 1-10.
120. Ward JM, Diwan BA, Ohshima M, Hu H, Schuller HM, Rice JM. Tumor-initiating and promoting activities of di(2-ethylhexyl) phthalate in vivo and in vitro. *Environ Health Perspect* 65 (1986) 279-291.
121. Weast RC, Astle MJ, Beyer WH (Eds). *CRC Handbook of Chemistry and Physics*, 65th ed. Boca Raton, CRC Press Inc, Florida 1984.
122. White RD, Carter DE, Earnest D, Mueller J. Absorption and metabolism of 3 phthalate esters by the rat small intestine. *Fd Cosmet Toxicol* 18 (1980) 383-386.
123. Wolowsky-Tyl R, Jones-Price C, Marr MC, Kimmel CA. Teratologic evaluation of diethylhexyl phthalate (DEHP) in Fischer 344 rats. *Teratology* 27 (1983) 85A.
124. Wolowsky-Tyl R, Jones-Price C, Marr MC, Kimmel CA. Teratologic evaluation of diethylhexyl phthalate (DEHP) in CD-1 mice. *Teratology* 27 (1983) 84A-85A.
125. Woodward KN, Smith AM, Mariscotti SP, Tomlinson NJ. Review of the toxicity of the esters of o-phthalic acid (phthalate esters), *Toxicity Review* 14, 183 p. Health and Safety Executive, London 1986.

126. Yagi Y, Nakamura Y, Tomita I, Tsuchikawa K, Shimoi N. Teratogenic potential of di- and mono-(2-ethylhexyl) phthalate in mice. J Environ Pathol Toxicol 4 (1980) 533-544.
127. Yanagita T, Kobayashi K, Enomoto N. Accumulation of hepatic phospholipids in rats fed di(2-ethylhexyl) phthalate. Biochem Pharmacol 27 (1978) 2283-2288.
128. Yanagita T, Kuzuhara S, Enomoto N, Shimad T, Sugano M. Effects of di(2-ethylhexyl) phthalate on the content of hepatic mitochondrial and microsomal phospholipids in the rat. Biochem Pharmacol 28 (1979) 3115-3121.
129. Yoon JS, Mason JM, Valencia R, Woodruff RC, Zimmering S. Chemical mutagenesis testing in Drosophila. IV. Results of 45 coded compounds tested for the National Toxicology Program. Environ Mutagen 7 (1985) 349-367.
130. Yoshikawa K, Tanaka A, Yamaha T, Kurata H. Mutagenicity study of nine monoalkyl phthalates and a dialkyl phthalate using Salmonella typhimurium and Escherichia coli. Fd Chem Toxicol 21 (1983) 221-223.
131. Zeiger E, Haworth S, Mortelmans K, Speck W. Mutagenicity testing of di(2-ethylhexyl) phthalate and related chemicals in Salmonella. Environ Mutagen 7 (1985) 213-232.